

**COMPARISON OF ZOOPLANKTON COMMUNITY STRUCTURES IN THREE
POLYCULTURE MODELS OF PONDS STOCKING *CTENOPHARYNGODON
IDELLUS***

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ABSTRACT: *The Zooplanktons community structures in three polyculture models of ponds, stocking mainly grass carp (*Ctenopharyngodon idellus*) were enumerated. We focused on the variability of a zooplankton community structure and compared its differences in the three polyculture models. Environmental factors and zooplankton abundance differed significantly among models. Nevertheless, model 2 had the minimum biomass of crustacean and zooplankton. Statistical analysis revealed that, NO_2^- -N, COD, NO_3^- -N, T, DO, PO_4^{3-} -P and pH are the most important environmental variables acting on the zooplankton assemblage. The differences in community structure of zooplankton and the growth rate of fish in the three Polyculture models showed that the model 2 was the most reasonable and efficient culture method, and Paddlefish (*Polyodon spathula*) can make more valuable contributions to market demand than bighead carp (*Aristichthys nobilis*).*

KEYWORDS: Zooplankton Community Structure, Pond Polyculture Model, Environmental Factors, *Ctenopharyngodon Idellus*

INTRODUCTION

Polyculture is one of the most important models of pond aquaculture system in China, in which, *Cyprinidae* fish is the main stocking species. Pond fish culture accounted for 70.6% of the freshwater aquaculture production in China (Fisheries statistic in China 2013). Though, China is the leading pond aquaculture production, but the problems of the unreasonable introduction of exotic species (Specify species, fish or other) and use of aquaculture drug become increasingly prominent during the culture period. These problems have restricted the further development of pond culture system. However, deformity of the ecological structure, blocking of material circulation and serious pollution of fresh water is the existing problems in aquaculture production system in China.

Zooplankton plays an important role in the water ecosystem structure and function. As a consumer of primary producers, it regulates the abundance of phytoplankton through predation, and as food of planktivorous fish. And its abundance directly impacts the resources of fish (Boudreau and Dickie 1992). Distribution of zooplankton is influenced by abiotic limitations. Some zooplankton species are limited by temperature, dissolved oxygen, salinity and other physicochemical factors; this may depends on the background of local environmental conditions. Compared with the large water area, such as reservoirs and lakes, the habitats of zooplankton communities in individual ponds may be different, due to the limited geographic range of certain species and chance events associated with colonization. Zooplankton community structure is influenced by many factors, such as predation, competition, but predation is the most important factor for variation of zooplankton assemblage. Some species

have even been shown to have different environmental preferences in different regions (Patalas 1971).

Paddlefish (*Polydon spathula*) was introduced from the United States 20 years ago. Early investigations indicated that paddlefish is primarily planktivores and insectivores (Kofoid 1900; Eddy and Simmer 1929). Many studies had confirmed that paddlefish was an indiscriminate filter feeder on zooplankton, especially cladoceran (Rosen and Hales 1981; Michaletz et al. 1982). Paddlefish are suitable to be stocked in large areas of reservoirs and lakes. But in ponds, there are few reports about the model of using paddlefish as Polyculture fish in China. In Chen et al. (2011) replaced bighead carp (*Aristichthys nobilis*) with paddlefish in polyculture ponds to compare the variety of water quality factors (chlorophyll a, transparency, total ammonia nitrogen, total nitrogen, and total phosphorus). And also, replacing bighead carps with paddlefish in ponds of mainly stocking grass carp. This study was designed to explore the relationship between zooplankton community and environmental factors, and to compare models that is most suitable for Polyculture system.

MATERIAL AND METHODS

Study Area

The study was conducted from June to October in 2010 in nine ponds near the “Shaotanhe” reservoirs. They are located in Hubei Province, China (30°50'N, 114°58'E). The nine ponds have the same area (110 m²) and similar depth (1.2 ± 0.2 m, mean ± SD). The water quality in “Shaotanhe” Reservoir is fresh and meets the water quality criteria in aquaculture.

Experimental Design

The experiment was designed into three polyculture models, stocking mainly, grass carp with few silver carp (*Hypophthalmichthys molitrix*), bighead carp, crucian carp (*Carassius auratus*) and paddlefish. The nine reconstructed ponds were divided into three groups, with triplicates in each group. Group one (model 1) including 250 Grass Carp, 35 Silver Carp, 40 Big head Fish, and 15 Goldfish, group two (model 2) 250 Grass Carp, 35 Silver Carp, 20 Big head Fish, 20 Paddlefish, and 15 Goldfish, and group three (model 3) 250 Grass Carp, 35 Silver Carp, and 15 Goldfish respectively. The individual weight of Grass Carp, Silver Carp, Big head Fish, Goldfish and Paddlefish was 87.31±9.19 g, 63.66±7.30 g, 55.01±3.89 g, 63.66±7.30 g and 150.15±11.9 g, respectively. The environmental factors of the three models were measured throughout the experimental period (Table 5). The experiment was conducted during the aquaculture production period in central China.

Pond Preparation and Fish Stocking

Prior to the trial, ponds were renovated with the same silt substrate. Aquatic vegetation, fish and other larger aquatic organisms were eradicated. Ponds were subsequently treated with lime (CaO, 250 kg ha⁻¹) and filled with water from “Shaotanhe” Reservoir. During the cultural period (June-October), 1/3 of the pond water was changed weekly. Ponds were unfertilized no longer during the entire experimental period.

On day 12, fingerlings collected from “Shaotanhe” Reservoir aquaculture farm; were released into the ponds. The fish were fed at a daily commercial floating juveniles grass carp feed to

satiation during the culture period. All of the fish were harvested by fishing net in December. The daily weight and net weight of paddlefish and bighead carp were calculated from the correlation formula (Chen et al. 2012).

Sampling of Zooplankton

Zooplankton samples were collected monthly (Jun 22, July 24, Aug 23 and Sep 23) at 08:30-10:30 in the morning. Horizontal hauls sampling (for qualitative samples), were taken, using a 64- μm mesh zooplankton net on the surface of water.

Laboratory Analysis

The collected samples were preserved in 5% buffered formaldehyde solution. Observation and identification of zooplankton to species level were done with the aid of an Olympus model microscope (magnification $\times 400$) model 230485 (Olympus America Inc., Center Valley, PA, USA) in the zooplankton laboratory. All cladoceran, Copepoda, rotifers and protozoan were identified to species using keys of Chiang and Du (1979), Sheng (1979), Shen (1990) and Wang (1965).

Rotifers and protozoan quantitative samples were obtained by filtering 1L water through a plankton net (mesh size 64 μm) and preserved in 5% buffered formaldehyde solution. The samples subsided for 48hrs in glass columns; then the supernatant was removed carefully and concentrated to 30 ml. Rotifer and protozoan samples were counted using 1 ml and 0.1 ml Sedgwick-R after counting chamber, respectively.

Crustacean and zooplankton samples were taken monthly. At each point, 10L mixed water, collected at two depths (0.5 m below the surface and 0.5 m away from the bottom), was filtered by a plankton net (mesh size 64 μm) and preserved with 5% formalin. Crustacean and zooplankton samples were all counted using a 1ml Sedgwick-R after counting chamber. Biomass of crustaceans were calculated from regression equations (Huang et al. 1984) of wet body weight and body length. Length measurements of cladoceran were made from the top of the head (helmet of *Daphnia* excluded) to the base of the shell spine; in Copepoda (copepodite and adult), from the top of the head to the end of the caudal rami (Yang et al. 1999). Average body length of crustacean was calculated from 20 animals. Weight of sample was estimated to be 0.003 mg. Samples were collected from mid-layers of the ponds so as to determine physical-chemical parameter using a 2.5 L organic glass hydrosphere, and preserved with chloroform.

Dissolved oxygen (DO), water temperature (T) and pH were measured in HQd Meters and IntelliCAL™ Probes (Hach Co., Loveland, CO, USA). Transparency (SD) was measured with a Secchi disc. All the water quality parameters, including total phosphorus (TP), phosphate ($\text{PO}_4^{3-}\text{-P}$), Ammoniacal nitrogen ($\text{NH}_3\text{-N}$), nitrite nitrogen ($\text{NO}_2^-\text{-N}$), nitrate nitrogen ($\text{NO}_3^-\text{-N}$), total nitrogen (TN), chemical oxygen demand (COD) and chlorophyll a (Chl-a) were measured within 24hrs in the hydration laboratory according to following standard methods (Greenberg et al. 1992; National Environmental Protection Bureau 2002).

Statistical Analysis

The software packages IBM SPSS Statistics 19 and CANOCO version 4.5 (Braak and Smilauer 2002) were used for statistical analyses. Redundancy analysis (RDA) was widely used to explain the relationship between the zooplankton communities and environmental factors (Marta 2012).

The variation of zooplankton density and biomass in different model ponds were analysed by SPSS 19.0. Diversity of zooplankton community was assessed using Simpson diversity index (Simpson 1949), Shannon-Wiener diversity index (Shannon et al. 1949) and Pielou index (Pielou 1975).

The dominant species of zooplankton were calculated according to Xu (1989), the computational formula is, where, N_i means the abundance of the species I , I means the occurrence frequency of this species, and N indicates the total abundance of zooplankton. A species are confirmed as the dominant species when the value of $Y \geq 0.02$.

The relationship between zooplankton and environmental factors and sampling Units

Redundancy analysis (RDA) was used to describe the relationships between the zooplankton abundance and the environmental variables (CANOCO program; Ter Braak and Smilauer 2002). RDA was chosen because the length of the gradient of the first DCA (Detrended Correspondence Analysis) axis running on zooplankton density data was 0.163; therefore, we used linear ordination techniques (Ter Braak and Prentice 1988). All zooplankton density data were $\log(x+1)$ transformed. Environmental factors (TP, $\text{PO}_4^{3-}\text{-P}$, $\text{NH}_4^+\text{-N}$, $\text{NO}_2^-\text{-N}$, $\text{NO}_3^-\text{-N}$, TN, COD, DO, pH, T and Chl-a) were selected by the automatic forward selection with 499 unrestricted Monte Carlo permutations. Bi-plot of relationship between environmental factors and zooplankton abundance (Fig. 3) And relationship between sampling units and zooplankton abundance (Fig. 4) Were presented using Cano Draw of CANOCO 4.5 (ter Braak and Smilauer 2002).

RESULTS

Zooplankton Community Composition

The dominant species of zooplankton in the three models were almost similar, which were absolutely dominated by protozoa and rotifer. *Urotricha factual* and *Lagynophrya conifers* were the dominant species of protozoan, rotifers were dominated by *Trichocerca possible*, Cladocera were dominated by *Moina micrura* and Copepoda dominated by *Thermocyclops holiness*. The zooplankton assemblage consisted of 133 genera (233 species), among which protozoan represented the highest number of species (154), approximately 66.5% of the total species of zooplankton, rotifers (60), which accounted for 25.4%. Cladocera (15) and Copepoda (4) accounted for 6.4% and 1.7%, respectively.

Density and biomass

Protozoa dominated the zooplankton assemblage in the three different models, with the highest density and biomass, followed rotifer the second dominant taxa, Cladocera and Copepoda accounted for only a small proportion (Table 1 and Table 2). The highest biomass of the total zooplankton occurred in model 1 ($39.19 \text{ mg}\cdot\text{L}^{-1}$), while model 2 and model 3 were very similar, their biomass were $24.70 \text{ mg}\cdot\text{L}^{-1}$ and $23.32 \text{ mg}\cdot\text{L}^{-1}$, respectively (Table1). The highest density and biomass of microzooplankton (Protozoa and Rotifer) was model 1 ($5.78 \times 10^5 \text{ Ind}\cdot\text{L}^{-1}$, $37.84 \text{ mg}\cdot\text{L}^{-1}$), while model 2 ($4.09 \times 10^5 \text{ Ind}\cdot\text{L}^{-1}$, $22.85 \text{ mg}\cdot\text{L}^{-1}$) and model 3 ($4.02 \times 10^5 \text{ Ind}\cdot\text{L}^{-1}$, $22.12 \text{ mg}\cdot\text{L}^{-1}$) had similar value.

Model 3 had the highest biomass ($2.57 \text{ mg}\cdot\text{L}^{-1}$) of macrozooplankton (Cladocera and Copepoda), while model 2 showed the least value ($0.54 \text{ mg}\cdot\text{L}^{-1}$). There was significant difference of macrozooplankton biomass between model 3 and model 2 ($P<0.05$), also between model 3 and model 1 ($P<0.05$).

Biodiversity Variation of Zooplankton

The results of the three diversity indices indicated that the biodiversity of zooplankton gradually reduced in the sequence of model 1, model 2 and model 3. The diversity indices of the three models showed a similar trend (Table 3).

Growth Parameters of Bighead Carp and Paddlefish

The parameters showed that the daily weight and net weight of bighead carp in model 2 significantly higher than model 1 ($P<0.05$); in model 2, the daily weight and net weight of paddlefish significantly higher than bighead carp (Table 4).

Water Quality Parameters

Water temperature and pH showed a similar temporal trend in the three model ponds. The level of DO was low during the whole experiment. The concentrations of TP, $\text{PO}_4^{3-}\text{-P}$ and TN were low in model 2, and there was a significant difference between model 1 and model 2 ($P<0.05$), model 3 and model 2 ($P<0.05$), there was no significant difference between model 1 and model 3. $\text{NH}_4^+\text{-N}$, $\text{NO}_3^-\text{-N}$ and COD decreased in order of sequence; model 1, 2 and 3. $\text{NO}_2^-\text{-N}$ was significantly high in June, and its concentration of model 1 and model 3 was lower than model 2.

Relationship between Zooplankton and Environmental Factors

In the current study Eigenvalues of the first and second axis were 0.794 and 0.156, respectively, and the two axes accounted for 92.1% of the variance in the zooplankton-environmental relationship (Table 6). Automatic forward selection indicated that the most significant variable of the 11 environmental variables were $\text{NO}_2^-\text{-N}$, which explained 36% of the total variance, followed by COD (28%), $\text{NO}_3^-\text{-N}$ (10%), T (8%), DO (7%), $\text{PO}_4^{3-}\text{-P}$ (5%), pH (4%). Axis 1 was positively related to TN (the correlation coefficient was 0.6097) and negatively related to $\text{NO}_2^-\text{-N}$ (-0.6409). Axis 2 was positively related to TP (0.6961) and $\text{NH}_4^+\text{-N}$ (0.6217), it was negatively related to DO (-0.8576) (Table 7).

As it is indicated on Fig.3, the protozoa and rotifer showed a positive relationship to COD, $\text{NH}_4^+\text{-N}$ and TP but were negatively related to DO and T. Cladocera was positively related to $\text{PO}_4^{3-}\text{-P}$, TN and TP but negatively related to $\text{NO}_2^-\text{-N}$ and DO, while Copepoda showed a positive relationship with T, TN and $\text{PO}_4^{3-}\text{-P}$ but was negatively related to $\text{NO}_2^-\text{-N}$ and COD. (Fig. 4) Showed the relationship between the four taxa of zooplankton and sampling units.

This bi-plot showed that the nine culture ponds were divided into three groups, the first group, including ponds 1, 2 and 3, which belongs to model 1 in this study. The second group, including pond 4, 5, 6 and 7 belongs to model 2 (except pond 7), Ponds 8 and 9 were the third group belonging to model 3. The largest abundance of protozoa and rotifer was the first group, followed by the second and third group; the second group had the least abundance of Cladocera, while the third group had the largest abundance of Cladocera and Copepoda.

DISCUSSION

Variation of zooplankton community structure and composition in three model culture ponds

The result of GU (1994), GU and Liu (1996) indicated that the average biomass of zooplankton in ponds of high production is about 10-20 mg/L in China, and microzooplankton (protozoa and rotifer) always dominated in these ponds, while macrozooplankton (Cladocera and Copepoda) is relatively low. It has been demonstrated that zooplanktivorous fish selectively prey on larger and more conspicuous zooplankton, mainly the large cladoceran (Elhigzi et al. 1995). Yang et al. (2011) also reported, zooplanktons were dominated by protozoa and rotifer in polyculture pond of grass carp, silver carp and common carp (*Cyprinus Carpio*). All these results of the various studies were similar to this research. Both silver carp and bighead carp can prey zooplankton (Cremer and Smitherman 1980), but the feeding capacity of bighead carp for zooplankton was greater than silver carp (Dong and Li 1995). Bighead carp has great water absorbing capacity and it can prey on Cladocera and Copepoda efficiently.

Yang and Huang (1992) studied the influence of silver carp and bighead carp on zooplankton community structure using the enclosure method, they found that in the enclosures where silver carp and bighead carp were stocked, the protozoa and some small-sized rotifers were dominated, while crustacean and large-sized rotifers were less. Similarly, Arcifa et al. (1986) have the same conclusion in their enclosure experiment.

In the present study, small-sized rotifers, such as *Trichocerca possible* and *Trichocerca gracilis* and protozoa were the dominant species in all the ponds. The activities of fish also influence the zooplankton community composition. The activity of crucian carp can make the resting eggs of rotifer hatched, increases the biomass of rotifer (Liu 1992). Macrozooplankton consume a broader size spectrum of particles and have a lower threshold food density below which they cannot maintain metabolic demand (Burns 1968). Some cyclopoids are omnivorous and prey on protozoans and rotifers. So the fish culture ponds, the number of small species was much greater than zooplankton.

Comparing with the average biomass of macrozooplankton (Cladocera and Copepoda) in the three model culture ponds, we found that model 3 had the largest value (10.40%), because there were no bighead carp and paddlefish in model 3. The biomass of Cladocera in model 2 was significantly lower than in model 1 and 3; due to a replacement of 6% bighead carp from model 1 to model 2. Paddlefish fed on zooplankton, especially Cladocera, it also preyed on Chironomidae larvae, juvenile fish and prawn, and its feeding habit, living habits and living space were similar with bighead carp (John and David 1986). Zhu et al. (2009) studied the food selection of paddlefish and pointed out that larval and juvenile paddlefish had the obvious food selection; they almost prey on cladoceran only, less copepods and aquatic insects, and barely prey on rotifers and phytoplankton. This shows paddlefish has great predation pressure on large zooplankton (Cladocera and Copepoda).

Relationship of environmental factors and zooplankton composition

The composition of zooplankton was influenced by physical, chemical, biological and geographic factors (MacArthur 1967). Olive and Pastuchova (2012) suggested that nutrients and dissolved oxygen were the most important physical-chemical factors for plankton diversity when they investigated the zooplankton communities with different traffic conditions in two

catchments. Karl et al. (2001) pointed out ponds with clear-water state was characterized by large cladoceran species and low chlorophyll-a concentration, unlike in the turbid-water ponds, high abundances of rotifers and inverse environmental conditions appeared.

In this present study, RDA indicated that most environmental factors were positively related to protozoa and rotifer, such as COD, $\text{NH}_4^+\text{-N}$, chlorophyll-a, $\text{NO}_3^-\text{-N}$ and TP, and these variables were all had the highest concentrations in model 1. As the flood of zooplankton, phytoplankton biomass is closely related to zooplankton biomass, and the concentration of chlorophyll-a can reflect the phytoplankton biomass. The bi-plot of the RDA indicated that ammonium showed the closest relationship with chlorophyll-a. Due to the favourable energy requirement, ammonium utilization is preferred by algae (Ward and Wetzel 1980), therefore, a slight increasing tendency of phytoplankton biomass can be expected even in the eutrophic-hypertrophic range. So ammonium influenced zooplankton indirectly. Guevara et al. (2009) also suggested that the enhancement of eutrophication could possibly increase rotifer abundance since they are able to thrive in the traffic environment.

In the current study, model 1 has the highest concentration of nutrients and highest abundance of protozoa and rotifer. This was consistent with the conclusion that with the increasing of nutrient concentration and eutrophication, small-sized species dominated in the zooplankton community (Chen et al. 2012), while Copepoda was highly negatively related to nutrient concentrations (Tadesse 2011). The bi-plot of Redundancy Analysis for zooplankton and environmental variables (Fig. 3) also showed similar results.

Feasibility analysis of polyculture of grass carp, silver carp, bighead carp, crucian carp and paddlefish

Grass carp, silver carp, bighead carp, crucian carp and paddlefish move at different water layers and feed on different preys in polyculture ponds; they enable a more efficient utilization of pond resources. Shi (2013) studied that during the period of juvenile, the feed coefficient of bighead carp is significantly higher than paddlefish, this indicates that paddlefish has higher utilization rate and conversion efficiency on zooplankton than bighead carp. GU et al. (2012) compared the benthic macroinvertebrates in three culture model stocking grass carp ponds, and deduced that the paddlefish can make valuable contributions to the biodiversity and ecosystem stability of fish ponds. Paddlefish has high value of ornament, nutrition and medicine (Chen 2011), and Chen (2011) discussed the value of replacing bighead carp with paddlefish, he concluded that paddlefish has higher economic value and social value than bighead carp. In this study, model 2 has lower biomass of zooplankton, especially macrozooplankton than model 1, which indicated that paddlefish can make higher utilization of zooplankton than bighead carp. In addition, the daily weight and net weight of paddlefish were significantly higher than bighead carp in model 2, and the market value of paddlefish is much higher than bighead carp in the Chinese market today, therefore, paddlefish more accords with the demand of people in China, and replacing bighead carp with paddlefish in polyculture pond is feasible.

CONCLUSIONS

The microzooplankton (protozoa and rotifer) dominated in the polyculture ponds, and associated with most of nutrients in the water, while macrozooplankton (Cladocera and Copepoda) were less in the three culture models, among which model 2 has the least biomass

of Cladocera and Copepoda. In addition, model 2 is the best aquaculture model of the three models. From this, we concluded that the first and second group (model 1 and model 2) was dominated by microzooplankton (protozoa and rotifer), while the third group had the largest abundance of macrozooplankton (Cladocera and Copepoda) compared to the former two groups. To confirm suitable combinations of fish species stocked in polyculture ponds, the factors of the utilization of fish for space resource, water, environment, the stability of ponds and market requirements should be taken into consideration.

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REFERENCES

- Arcifa, M. S., Northcote, T. G. and Froehlich, O. (1986) Fish-zooplankton interactions and their effects on water quality of a tropical Brazilian reservoir, *Hydrobiologia*, 139 49–58.
- Boudreau, P. R. and Dickie, L. M. (1992) Biomass spectra of aquatic ecosystem in relation to fisheries yield, *Canadian Journal of Fisheries and Aquatic Sciences*, 49 1528-1538.
- Burns, C. W. (1968) The relationship between body-size of filter-feeding zooplankton and the maximum size of particle ingested, *Limnology and Oceanography*, 13 675-678.
- Chen, F. Z., Ye, J. L., Shu, T. T., Sun, Y. and Li, J. (2012) Zooplankton response to the lake restoration in the drinking-water source in Meiliang Bay of subtropical eutrophic Lake Taihu, China, *Limnologica*, 42 189-196.
- Chen, J. W., Zhu, Y. J., Zhao, J. H., Feng, X.B. and Yang, D. G. (2011) Water Quality and N & P Utilization in Polyculture Ponds with Paddlefish *Polydon spathula*, *Hubei Agricultural Sciences*, 50 1434-1438.
- Chen, J., Xiong, B. X., Zhu, Y. T., Gu, Q. H., Huang, J. and Wang, Q. (2012) Comparative on Fish Growth and Efficiency of Three Polyculture Models in Grass Carp Mainly Culture Pond, *Journal of Hydroecology*, 33 108-112.
- Chen, J. Y. (2011) Discussion of value of replacing big head carp with paddlefish, *Current Fisheries*, 6 62-66.
- Chiang, S. C. and Du, N. S. (1979) *Fauna Sinica, Crustacea, Freshwater Cladocera*. Science Press, Academia Sinica, Beijing.
- Cottenie, K., Nuytten, N., Michels, E. and Meester, L. D. (2001) Zooplankton community structure and environmental conditions in a set of interconnected ponds, *Hydrobiologia*, 442 339-350.
- Cremer, M. C. and Smitherman, R. O. (1980) Food habits and growth of silver and bighead carp in cages and ponds, *Aquaculture*, 20 57-64.
- Dong, S. L. and Li, D. S. (1995) Comparative studies on the feeding capacity of silver carp (*Hypophthalmichthys molitrix*) and bighead carp (*Aristichthys nobilis*), *Oceanologia et Limnologia Sinica*, 26 53-57.
- Eddy, S. and Simmer, P. A. (1929) Notes on the food of paddlefish, *Transactions of the Illinois State Academy of science*, 21: 59-68.
- Elhigzi, F. A., Haider, R. and Larsson, S. A. P. (1995) Interactions between Nile tilapia (*Oreochromis niloticus*) and cladocerans in ponds (Khartoum, Sudan), *Hydrobiologia*, 307 263-272.

- Greenberg, A. E., Clesceri, L. S. and Eaton, A. D. (1992) Standard Methods for the Examination of Water and Wastewater, 18th ed. American Public Health Association, Washington, DC.
- Gu, Q. H, Xiong, B. X., Zhu, Y. T., Yang, X. F. and Shi, P. S. (2012) Comparison of benthic macroinvertebrates in three polyculture models of ponds stocking mainly *Ctenopharyngodon idellus*, *Desalin Water Treat*, 45 26-39.
- Gu, X. H. (1994) Preliminary analysis on the community structure of the plankton in different types of fish ponds, *Journal of Lake Sciences*, 6 276-282.
- Gu, X. H. and Liu, G. Y. (1996) Impact of silver carp and big head carp in fish ponds on plankton, *Rural Eco-Environment*, 12 6-10.
- Guevara, G., Lozano, P. and Reinoso, G. (2009) Horizontal and seasonal patterns of tropical zooplankton from eutrophic Prado reservoir (Colombia), *Limnologia*, 39 128–139.
- Huang, X. F., Chen, X. M., Wu, Z. T. and Hu, C. Y. (1984) Studies on changes in abundance and biomass of zooplankton in Lake Donghu, Wuhan, *Acta Hydrobiologica Sinica*, 8 345-358.
- Illyova, M. and Pastuchova, Z. (2012) The zooplankton communities of small water reservoirs with different trophic conditions in two catchments in western Slovakia, *Limnologia*, 42 271-281.
- John, S. B. and David, R. B. (1986) Impact of paddlefish on plankton and Water Quality of Catfish Ponds, *The Progressive Fish-Culturist*, 48 177-183.
- Karl, C., Nele, N., Erik, M. and Luc, D. M. (2001) Zooplankton community structure and environmental conditions in a set of interconnected ponds, *Hydrobiologia*, 442 339–350.
- Kofoed, C. A. (1900) Notes on the natural history of *Polyodon*. *Science* 11 252.
- Liu, Q., Zhao, Y. B., Wang, Y. and Li, Y. H. (1992) Investigations on plankton of high yield fish ponds in Zhenlai district, Jilin Prov, *Journal of Dalian Fishes College*, 6 14-27.
- MacArthur, R. H. and Wilson, R. O. (1967) *The Theory of Island Biogeography*. Princeton University Press. Princeton, New Jersey.
- Michaletz, P. H., Rabeni, C. F., Taylor, W. W. and Russell, T. R. (1982) Feeding ecology and growth of young of the year paddlefish in hatchery ponds, *Transactions of the American Fisheries Society*, 111 700-709.
- National Environmental Protection Bureau (2002) *Guide of Analytical Method on Monitoring Water and Waste Water*. Environmental Science Press, Beijing, pp. 1-199.
- Patalas, K. (1971) Crustacean plankton communities in forty-five lakes in the Experimental Lakes Area, northwestern Ontario, *Journal of the Fisheries Research Board of Canada*, 28 231-244.
- Pielou, E. C. (1975) *Ecological Diversity*. Wiley, New York.
- Rudolph, A. R. and Donald, C. H. (1981) Feeding of paddlefish, *Polyodon spathula*. *COPEIA*, 2 441-455.
- Shannon, C. E. and Weaver, W. (1949) *The Mathematical Theory of Communication*. Urbana, pp. 125.
- Shen, Y. F., Zhang, Z. S., Gong, X. J., Gu, M. R., Shi, Z. X. and Wei, Y. X. (1990) *Modern Biomonitoring Techniques using Freshwater Microbiota*. Beijing, China.
- Sheng, J. R. (1979) *Fauna Sinica, Crustacea, Freshwater Copepoda*. Beijing, China.
- Shi, P. S. (2013) Cloning and expression of fatty acid synthase gene and comparison on growth, muscle quality of *Polyodon spathula* and *Aristichthys nobilis*. Wuhan, Huazhong Agricultural University.
- Simpson, E. H. (1949) Measurement of species diversity, *Nature*, 163 688.

- Tadesse, F., Seyoum, M., Michael, S. (2011) Zooplankton community structure and ecology of the tropical-highland Lake Hayq, Ethiopia, *Limnologica*, 41 389-397.
- Ter Braak, C. J. F. and Prentice, I. C. (1988) A theory of gradient analysis, *Advances in Ecological research*, 18 93-138.
- Ter Braak, C. J. F. And Smilauer, P. (2002) CANOCO reference manual and CanocoDraw for Windows User's guide: Software for Canonical Community Ordination (version 4.5). Microcomputer Power, Ithaca, NY, US.
- Wang, J. J. (1965) *Fauna Sinica, Rotifer*. Beijing, China .
- Ward, A. K. and Wetzel, R. G. (1980) Interactions of light and nitrogen source among planktonic blue-green algae, *Archiv für Hydrobiologie*, 90 1–25.
- Xu, Z. L. and Chen, Y. Q. (1989) Aggregated intensity of dominant species of zooplankton in autumn in the East China Sea and Yellow Sea, *Journal of Ecology*, 8 13-15.
- Yang, J. L., Gao, Q. F., Dong, S. L., Wang, F. and Tian, X. L. (2011) Effect of silver carp culture on the composition of plankton and suspended particulate matter in polyculture ponds, *Periodical of Ocean University of China*, 41 23-29.
- Yang, Y., Huang, X. F. and Liu, J. K. (1999) Long-term changes in crustacean zooplankton and water quality in a shallow, eutrophic Chinese lake densely stocked with fish, *Hydrobiologia*, 391 195-203.
- Yang, Y. F. and Huang, X. F. (1992) The influence of silver carp and bighead on the zooplankton community structure, *Journal of Lake Sciences*, 4 78-86.
- Zhu, A. M., Liang, Y. Q., Huang, D. M., Hu, X. J., Wu, S. G. and Hu, C. L. (2009) Study on starting-feeding size and food selection of larval and juvenile paddlefish, *Acta Hydrobiologica Sinica*, 33 1202-1206.

ABSTRACT

Table 1. The change of different months of density (ind.·L⁻¹) and biomasses (mg·L⁻¹) of zooplankton in different culture models

Species	Month	Model I		Model II		Model III	
		Density	Biomass	Density	Biomass	Density	Biomass
Protozoa	6	0.80×10 ⁵	2.93	2.37×10 ⁵	7.66	2.83×10 ⁵	8.86
	7	6.49×10 ⁵	31.39	4.25×10 ⁵	15.41	3.52×10 ⁵	14.04
	8	8.70×10 ⁵	39.28	4.12×10 ⁵	18.25	3.58×10 ⁵	11.76
	9	6.64×10 ⁵	36.38	5.16×10 ⁵	18.73	5.76×10 ⁵	17.60
	Avg.	5.66×10 ⁵	27.50	3.98×10 ⁵	15.01	3.92×10 ⁵	13.07
Rotifer	6	1.19×10 ⁴	16.02	0.89×10 ⁴	10.10	0.27×10 ⁴	1.87
	7	1.20×10 ⁴	8.21	0.93×10 ⁴	4.27	1.44×10 ⁴	22.24
	8	1.40×10 ⁴	9.75	1.61×10 ⁴	11.04	0.69×10 ⁴	9.36
	9	0.99×10 ⁴	7.41	1.02×10 ⁴	5.92	1.44×10 ⁴	2.76
	Avg.	1.20×10 ⁴	10.35	1.12×10 ⁴	7.83	0.96×10 ⁴	9.06
Cladocera and Copepoda	6	212.03	3.79	78.75	0.77	174.10	2.66
	7	34.57	0.33	29.10	0.34	50.70	0.43
	8	33.57	0.37	25.45	0.36	342.40	3.99
	9	81.53	0.89	56.70	0.67	305.00	3.21
	Avg.	90.43	1.35	47.50	0.54	218.05	2.57
Total zooplankton	6	0.92×10 ⁵	22.74	2.46×10 ⁵	18.53	2.86×10 ⁵	13.39
	7	6.61×10 ⁵	39.93	4.34×10 ⁵	20.02	3.66×10 ⁵	36.71
	8	8.84×10 ⁵	49.40	4.28×10 ⁵	29.65	3.65×10 ⁵	25.11
	9	6.74×10 ⁵	44.68	5.26×10 ⁵	25.32	5.91×10 ⁵	23.57
	Avg.	5.78×10 ⁵	39.19	4.09×10 ⁵	23.38	4.02×10 ⁵	24.70

Table 2. The percentage of mean density and biomass of protozoa, rotifer, Cladocera and Copepoda among the total zooplankton

Models	Percentage of mean density from the total (%)					Percentage of mean biomass from the total (%)				
	Total ($\times 10^5 \cdot L^{-1}$)	Protozoa (%)	Rotifer (%)	Cladocera (%)	Copepoda (%)	Total ($mg \cdot L^{-1}$)	Protozoa (%)	Rotifer (%)	Cladocera (%)	Copepoda (%)
Model I	5.78	97.924	2.060	0.007	0.009	39.19	70.162	26.406	2.314	1.118
Model II	4.09	97.250	2.738	0.003	0.009	23.38	64.210	33.500	0.922	1.368
Model III	4.02	97.565	2.381	0.015	0.039	24.70	52.911	36.683	4.029	6.377

Table 3. Average diversity indices of zooplankton in each month in different culture models

Diversity Index	Models	Jun	Jul	Aug	Sep	Average
Simpson index	Model I	9.4395	10.0782	13.7885	16.6463	12.4881
	Model II	4.7070	11.5470	14.6039	15.3586	11.5541
	Model III	4.0273	10.2802	14.0140	17.1775	11.3748
Shannon-Wiener index	Model I	2.8864	2.9359	3.0898	3.2558	3.0420
	Model II	2.3052	2.8817	3.1759	3.1271	2.8725
	Model III	2.1425	2.8713	3.0389	3.1541	2.8017
Pielou index	Model I	0.4656	0.4613	0.5004	0.5238	0.4878
	Model II	0.3710	0.4703	0.5164	0.5102	0.4670
	Model III	0.3409	0.4582	0.4958	0.5112	0.4515

Table 4. Growth parameters of bighead carp and paddlefish in the three polyculture models

Growth parameters	Bighead carp		paddlefish
	Model 1	Model 2	Model 2
Initial weight (g)	55.01±3.89	55.01±3.89	150.09±6.88
Harvest weight (g)	140.33±8.95 ^a	176.83±13.72 ^b	323.56±11.10 ^c
Daily weight (g·d ⁻¹) ₁₎	0.71±0.05 ^a	1.02±0.09 ^b	1.45±0.05 ^c
Net weight (g)	85.32±5.87 ^a	121.82±10.71 ^b	173.47±6.36 ^c

Note: Values with different letters in the same row indicated significant differences ($P < 0.05$).

Quoted from Chen et al. (2012)

Table 5. The results of physico-chemical parameters of water in three model ponds

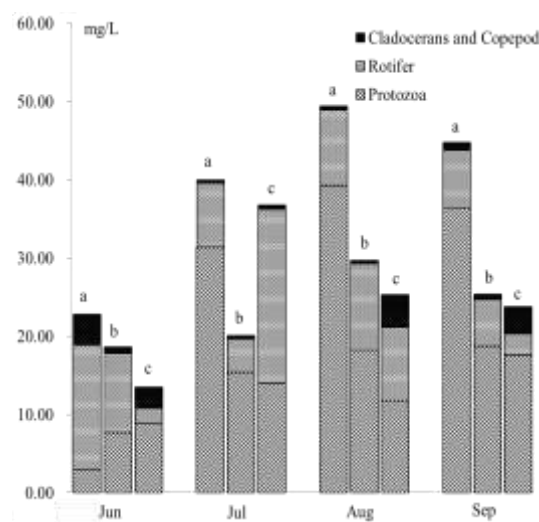
Models	Time	TP, mg.L ⁻¹	PO ₄ ³⁻ P, mg.L ⁻¹	NH ₄ ⁺ N, mg.L ⁻¹	NO ₂ ⁻ N, mg.L ⁻¹	NO ₃ ⁻ N, mg.L ⁻¹	TN, mg.L ⁻¹	COD, mg.L ⁻¹	DO, mg.L ⁻¹	pH	T, °C	Chlorophyll a, mg.L ⁻¹
Model I	Jun.	0.125	0.066	0.504	0.011	0.131	0.888	6.952	2.307	7.977	28.967	0.041
	Jul.	0.119	0.072	0.554	0.000	0.030	0.916	6.344	4.277	8.137	30.200	0.094
	Aug.	0.123	0.010	0.212	0.000	0.000	1.650	6.344	2.187	7.497	30.267	0.229
	Sep.	0.095	0.022	0.346	0.002	0.000	1.254	6.890	1.813	7.200	24.933	0.179
	Oct.	0.109	0.024	0.310	0.002	0.086	1.548	6.833	2.937	7.227	20.800	0.356
	Mean	0.114	0.039	0.385	0.003	0.049	1.251	6.672	2.704	7.607	27.033	0.180
Model II	Jun.	0.048	0.013	0.484	0.027	0.178	0.756	6.578	4.263	8.083	28.533	0.058
	Jul.	0.055	0.023	0.557	0.000	0.039	1.029	6.410	5.010	8.050	30.333	0.087
	Aug.	0.092	0.010	0.216	0.000	0.000	0.947	6.410	3.173	7.580	30.500	0.155
	Sep.	0.073	0.032	0.405	0.002	0.000	0.961	5.980	2.327	7.133	25.100	0.152
	Oct.	0.085	0.037	0.075	0.003	0.008	1.311	6.593	4.497	7.283	20.833	0.275
	Mean	0.071	0.023	0.347	0.007	0.045	1.001	6.394	3.854	7.626	27.060	0.145
Model III	Jun.	0.070	0.023	0.399	0.006	0.101	0.863	7.473	2.933	7.820	29.233	0.053
	Jul.	0.109	0.032	0.520	0.000	0.018	1.214	5.929	3.853	7.850	30.600	0.094
	Aug.	0.092	0.014	0.149	0.000	0.000	1.212	5.929	3.467	7.613	30.900	0.173
	Sep.	0.095	0.049	0.303	0.002	0.000	1.379	5.048	2.223	7.277	25.167	0.141
	Oct.	0.090	0.053	0.088	0.005	0.000	1.470	5.563	5.230	7.667	21.267	0.261
	Mean	0.091	0.034	0.292	0.003	0.024	1.228	5.989	3.541	7.645	27.433	0.144

Table 6. Summary statistics of RDA

Axes	1	2	3	4
Eigenvalues	0.794	0.156	0.041	0.009
Species-environmental correlations	1.000	1.000	1.000	1.000
Cumulative percentage variance of species data	79.4	95.0	99.1	100
Sum of all eigenvalues	1.000			
Sum of all canonical eigenvalues	1.000			

Table 7. Redundancy analysis Inter-set correlations of environmental variables with axes

Environment variable	Axis 1	Axis 2	Axis 3	Axis 4
TP, mg.L ⁻¹	0.3518	0.6961	-0.3731	0.3684
PO ₄ ³⁻ -P, mg.L ⁻¹	0.5402	0.4662	-0.2380	0.1882
NH ₄ ⁺ -N, mg.L ⁻¹	0.0254	0.6217	0.0627	0.5939
NO ₂ ⁻ -N, mg.L ⁻¹	-0.6409	-0.3914	0.5469	-0.2402
NO ₃ ⁻ -N, mg.L ⁻¹	-0.2508	0.4296	0.7235	-0.0149
TN, mg.L ⁻¹	0.6097	0.4333	-0.5291	0.1825
COD, mg.L ⁻¹	-0.2338	0.6179	-0.5728	-0.0566
DO, mg.L ⁻¹	-0.2703	-0.8576	0.2117	-0.2859
pH	0.0459	-0.3534	0.1139	-0.3320
T, °C	0.5052	-0.5516	-0.2860	0.3104
Chl-a, mg.L ⁻¹	0.0901	0.5767	-0.5265	0.2637

**Fig 1. The biomass of zooplankton in each month in different stocking models (a, b, c represent model I, II, III, respectively)**

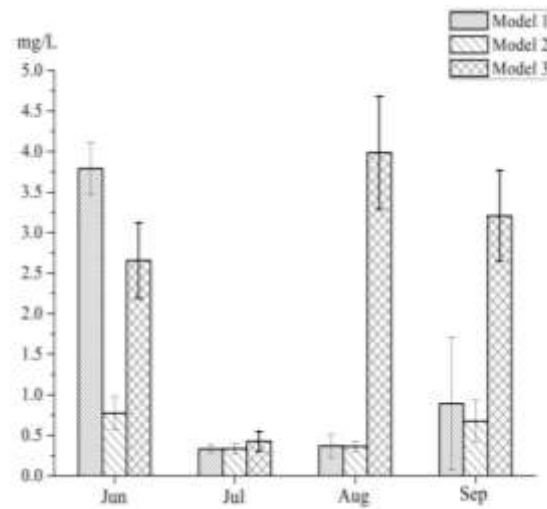


Fig 2.The biomass of crustacean in each month in different stocking models

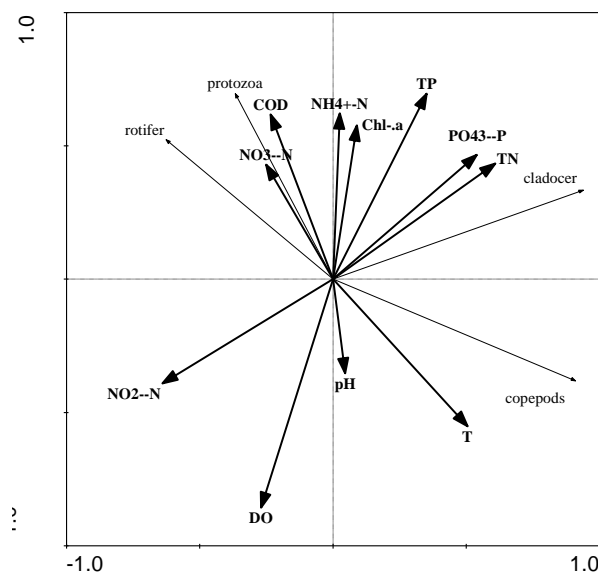


Fig 3. Bi-plot of the Redundancy Analysis for zooplankton abundance and environmental variables. Broken arrows represent taxa of zooplankton whereas bold arrows are environmental variables (rotifer-rotifer; protozoa-protozoa; cladocera-cladocera; copepods- copepods)

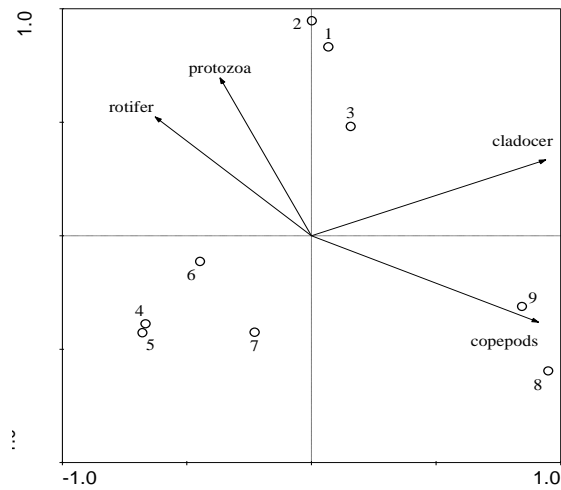


Fig 4. Bi-plot of the Redundancy Analysis for zooplankton abundance and sampling units (1, 2 and 3 were three ponds of model I , 4, 5 and 6 were three ponds of model II , 7, 8 and 9 were three ponds of model III)